



## High-Efficiency Adsorption and Sustainable Desorption of Anionic Dyes Using a Novel AC–Fe<sub>2</sub>O<sub>3</sub>/ZnO Nanocomposite

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### Keywords:

Activated carbon, Eosin Blue, Fuschin Acid, zinc oxide, organic pollutants, Recovery study

### ABSTRACT

This study proposes the preparation of a novel adsorbent material, AC-ZnO-Fe<sub>2</sub>O<sub>3</sub>, synthesized using activated carbon derived from plastic waste and metal oxides produced via co-precipitation methods. The structural and physicochemical characteristics of the prepared adsorbent were examined using FTIR, XRD, BET, and EDX analyses. Particle size estimation performed by SEM revealed values of 63.29 nm for activated carbon (AC), 91.69 nm for Fe<sub>2</sub>O<sub>3</sub>, 68.79 nm for ZnO, and 72.11 nm for the composite (AC-ZnO-Fe<sub>2</sub>O<sub>3</sub>). Zeta potential measurement recorded a negative surface charge of -42.2 mV. The batch adsorption technique was employed to evaluate the removal efficiency of two anionic dyes, Eosin Blue (EB) and Fuchsin Acid (FA), under varying operational conditions, including equilibrium time, temperature, pH, and agitation speed. Under optimal conditions, the removal efficiencies reached 77.17% and 88.21% for EB and FA, respectively. Adsorption isotherm analyses demonstrated that the Freundlich model provided the best results. In addition, thermodynamic parameters were calculated, and a negative  $\Delta G^{\circ}$  value was obtained, indicating a spontaneous adsorption process and confirming the endothermic nature of the reaction. Finally, Desorption experiments were conducted to achieve high recovery efficiencies of 98.46% For EB and 75.81% For FA, using water as the solvent.

## 1. Introduction

Environmental pollution has become one of the most pressing global challenges, threatening both ecosystems and human health [1]. In recent decades, rapid technological development and industrial expansion have intensified the release of contaminants into natural systems, thereby exacerbating environmental pollution [2-3]. Among the various forms of pollution, water contamination remains a critical issue because of its direct impact on life, agriculture, and industry [4]. Water pollution is defined as any alteration in the physical, chemical, or biological characteristics of water that renders it unsuitable for its intended use [5]. Pollutants present in aquatic systems can be broadly categorized into as inorganic pollutants, such as heavy metals, and organic pollutants, including dyes, hydrocarbons, and pesticides. Both pose severe environmental and health risks [6]. Synthetic dyes are among the most persistent and hazardous organic pollutants and are widely used in the textile, plastic, paper, cosmetics, rubber, and leather industries [7].

The inherent physicochemical stability of these compounds, coupled with their resistance to biodegradation, renders them particularly persistent in aquatic environments and difficult to remove using conventional treatment processes [8]. Their discharge into water bodies significantly reduces light penetration, thereby limiting photosynthetic activity and impairing primary productivity. This disruption subsequently contributes to a decline in dissolved oxygen concentrations, adversely affecting aquatic ecosystems. In addition, to these ecological impacts several studies have highlighted their potential to induced mutagenic and carcinogenic effects, posing serious risks to environmental and human health [9-11]. Among the wide spectrum of synthetic dyes, Eosin Blue (EB) and Fuchsin Acid (FA) have attracted particular concern due to their persistence and environmental impact. Eosin Blue is an anionic xanthene dye characterized by high aqueous solubility and widespread application in textile dyeing, ink production, and biological staining processes [12-13]. In contrast, Fuchsin Acid belongs to triphenylmethane class of anipnic dyes and is distinguished by a highly conjugated chromophoric structure that imparts strong color stability and resistance to degradation, thereby enhancing its environmental persistence [14-15]. Therefore, the development of cost-effective, efficient, and eco-friendly remediation technologies is essential for the sustainable treatment of dye-contaminated wastewater [16]. Among the several existing methods, adsorption stands out because of its high efficiency, operational simplicity, low cost, and environmental compatibility [17-19]. Activated carbon (AC) remains the most widely used adsorbent, offering a large surface area, high porosity, and chemical stability [20]. It is easily regenerated and produced from

sustainable materials such as fruit peels, coconut shells, and single-use plastic waste [21-22]. In recent years, nanomaterials have gained significant attention owing to their unique physicochemical, optical, and magnetic properties, primarily resulting from their nanoscale dimensions and high surface-to-volume ratio [23-24]. Metal oxidized nanoparticles such as Ag, ZnO, TiO<sub>2</sub>, and Fe<sub>2</sub>O<sub>3</sub> have demonstrated remarkable potential in water purification applications, while mixed metals oxides (e.g., Fe-Mn, Fe-Ti, Fe-Cr) exhibit synergistic effects that further enhance adsorption efficiency and stability [25-26]. This study aims to investigate the effectiveness of integrating activated carbon with metal oxide nanoparticles, offering a sustainable and highly effective approach for removing persistent dyes such as Eosin blue and Fuchsin Acid. This is achieved by combining the high surface area and adsorption capacity of activated carbon with the catalytic and reactive characteristics of nanostructured oxidized to achieve superior performance through synergistic interaction.

## 2. Materials and Methods

### 2.1. Chemicals

All chemicals used in this study were of analytical grade. Ammonium hydroxide (25%), NaOH and KOH were Obtained from Merck. Hydrochloric acid (37%) and nitric acid (65%) were supplied by GCC, and concentrated sulfuric acid (95%) was purchased from J.T. Baker. Ferric citrate[C<sub>6</sub>H<sub>5</sub>O<sub>7</sub>Fe.5H<sub>2</sub>O] and the dyes EB and FA were provided by BDH. Zinc acetate dihydrate [Zn (CH<sub>3</sub>COO)<sub>2</sub> .2H<sub>2</sub>O] was obtained from Riedel-De Haen Agseeleze Hannover, and ethanol (96%) from Fisher chemical.

### 2.2. Instrumentation

The functional groups on the surface of the prepared adsorbent were identified using FTIR in the range of 400–4000 cm<sup>-1</sup> with a Shimadzu FTIR -8400s Spectrophotometer. The maximum adsorption wavelength was determined using a Shimadzu UV-Vis 1800 spectrophotometer. The surface morphology and elemental composition were examined via SEM and EDX using a Vega A3 microscope at 20kV. The crystalline phase and crystallite size were analyzed by XRD using a Philips PW1370 diffractometer with Cu K $\alpha$  radiation ( $\lambda=1.54\text{\AA}$ ). Additionally, EDX analysis was performed using a Zeiss microscope to determine the elemental composition of the prepared adsorbent.

### 2.3. Synthesis of AC-ZnO-Fe<sub>2</sub>O<sub>3</sub> Nanocomposite

Activated carbon (AC) was prepared from plastic waste collected from different source (drinking water bottles made of polyethylene terephthalate). The waste was washed with deionized water, dried at 65 °C, crushed, and sieved (2 mm). A 25 g portion of the material was mixed with 25 mL of concentrated sulfuric acid (H<sub>2</sub>SO<sub>4</sub>) and stirred magnetically at 90 °C for 2h. The resulting black carbonaceous material was washed several times with deionized water and dried at 80 °C for 4h, followed by carbonization at 300 °C for 3h [22]. The sample was filtered and dried again at 80 °C for 5h. It was then oxidized with 32.5% nitric acid (HNO<sub>3</sub>) under continuous stirring at 60 °C for 5h, washed with deionized water until neutral pH, and dried at 80 °C the preparation method was depended [27].

Nanostructured metal oxidizes zinc oxide (ZnO), and ferric oxide (Fe<sub>2</sub>O<sub>3</sub>) were synthesized via the co-precipitation method . For ZnO nanoparticles were prepared by dissolving 2.19809 g of zinc acetate dihydrate in 300mL of distilled water under magnetic stirring. A KOH solution (1.2155 g in 50 mL distilled water) was added gradually until a white precipitate formed. The precipitate was then calcined at 400 °C for 180 min to obtain ZnO nanoparticles [29-30]. Similarly, Fe<sub>2</sub>O<sub>3</sub> nanoparticles, dissolved (1 g) of ferric citrate was dissolved in 100 mL of distilled water to form an orange solution. The pH was adjusted to neutrality using NaOH (1%), followed by the addition of ethanol to induce precipitation. The precipitate was separated by decantation, washed several times with ethanol, dried at 200 °C, and calcined at 300 °C [28].

Finally, the synthesized ZnO and Fe<sub>2</sub>O<sub>3</sub> nanoparticles were supported onto the activated carbon surface to obtain the AC-ZnO-Fe<sub>2</sub>O<sub>3</sub> nanocomposite. ZnO and Fe<sub>2</sub>O<sub>3</sub> nanoparticles (0.04 g each) were dispersed in 40mL of deionized water, followed by the addition of 0.2 g of activated carbon. The mixture was sonicated for 60 min to ensure uniform dispersion and then subjected to hydrothermal treatment at 180 °C for 4h. The resulting composite was separated, washed, and dried at 80 °C for 3hours to obtain the final AC-ZnO-Fe<sub>2</sub>O<sub>3</sub> nanocomposite [30].

### 2.4. Preparation standard solutions of Eosin Blue and Fuchsin Acid dyes

A standard stock solution with a concentration of 1000 mg/L was prepared, and various dye concentrations were subsequently prepared. The maximum wavelength of Eosin Blue (EB) and Fuchsin Acid (FA) was determined using 10 mg/L, and then measured using a UV-Vis spectrophotometer within a wavelength range of 200–800 nm. The results indicated that the maximum wavelengths were 518.5 nm and 546.5 nm for EB and FA, respectively.

## 2.5. Experiments of Adsorption

The adsorption process refers to the retention of ions, atoms, or molecules from liquids, gases, or dissolved solids onto the surface of a solid material. This phenomenon provides valuable insights into the nature of the adsorbed molecules, characteristics of the adsorbent surface, and interaction mechanisms between them [31]. In this study, adsorption experiments were conducted to remove anionic dyes from aqueous solutions and determine the optimal conditions for adsorption, including equilibrium time, pH, temperature, and mixing speed. These experiments were designed to evaluate the adsorption efficiency of the prepared adsorbent under controlled laboratory conditions. Each experiment was performed using 10mL of dyes solution with initial concentrations of 100 and 200 mg/L for Eosin blue and Fuchsin Acid, respectively.

A 0.01g portion of the previously prepared adsorbent was used, and the solutions were agitated at a speed of 150 rpm under neutral pH and room temperature to determine the equilibrium time for each dye. The removal percentage and adsorption capacity were calculated using the following equations [32]:

$$\% AD = [(C_0 - C_e) / C_0] \times 100$$

$$Q_e \text{ (mg/g)} = [V \times (C_0 - C_e)] / W$$

:Where

AD% is the adsorption ratio

C<sub>0</sub> is the initial dye concentration, (mg/L)

C<sub>e</sub> is the equilibrium dye concentration (mg/L)

Q<sub>e</sub> is the adsorption capacity (mg/g) V is the volume of the dye solution (L)

W is the mass of the adsorbent (g)

The adsorption isotherm experiments were conducted under different conditions from those in the optimization study. The initial dye concentrations and solution volumes were increased to obtain sufficient mass of adsorbate for the adsorbent dose, and thus to accurately estimate the equilibrium parameters. This resulted in a total amount of dyes greater than the adsorption capacity of the adsorbents, and thus the Langmuir Q<sub>max</sub> values fit the mass balance.

All experiments were performed in triplicate, and the reported values represent the mean ± standard deviation (SD).

## 2.6. Adsorption Isotherm

Adsorption isotherms describe the relationship between the amount of adsorbate retained on the surface of the adsorbent and its concentration in solution at a constant temperature (equilibrium state). Studying adsorption isotherms helps to understand the phenomenon associated with the adsorption process, the interaction between the adsorbent surface and dye molecules, and provides valuable insights into the thermodynamic behavior of adsorption systems. In the present study, two isotherm models were employed to describe the adsorption behavior: The Langmuir and Freundlich isotherms [33]. The Langmuir isotherm assumes that adsorption occurs on a homogeneous surface in a monolayer manner and that the process can be either physical or chemical, depending on the type of bonding between the adsorbate molecules and the adsorbent surface. The Langmuir model is expressed by the following equation [17]:

$$C_e/Q_e = (1 / (Q_{\max} \times K_L)) + (C_e/Q_{\max})$$

Where:

$Q_{\max}$ (mg/g) represents the maximum adsorption capacity for dye molecules at constant temperature

$K_L$  (L/mg) is the Langmuir constant related to the affinity of binding sites

on the other hand, The Freundlich isotherm, is generally applied to describe physical adsorption on heterogeneous surfaces, where multilayer adsorption occurs [34]. It is represented by the following equation:

$$\text{Log } Q_e = \text{Log } K_f + (\text{Log } C_e/n)$$

Where:

$K_f$  is the Freundlich constant indicating adsorption capacity

The value of the heterogeneity factor ( $n$ ) indicates the favorability of the adsorption process

When  $n > 1$ : the adsorption process is favorable and efficient (strong adsorption)

When  $0 < n < 1$ : the adsorption process is weak or unfavorable

When  $n = 0$ , the adsorption process is irreversible or does not occur.

## 2.7. Recovery

This study was conducted to recover the molecules of Eosin Blue and Fuchsin Acid dyes and returned them to the solution after completing the adsorption process under the selected optimum condition. The adsorbent material (0.01 g) and dye solution (10 mL) were taken, and

the mixture was separated using a centrifuge at 4500 rpm for 10 min. Water was used as a recovery solvent to break the bond formed between the dye molecules and the adsorbent surface molecules, and then it was liberated to solution once again and measured at the selected wavelength. The recovery percentage of the dyes was calculated using the following equation [35]:

$$\%R_e = [C_L / (C_0 - C_e)] \times 100$$

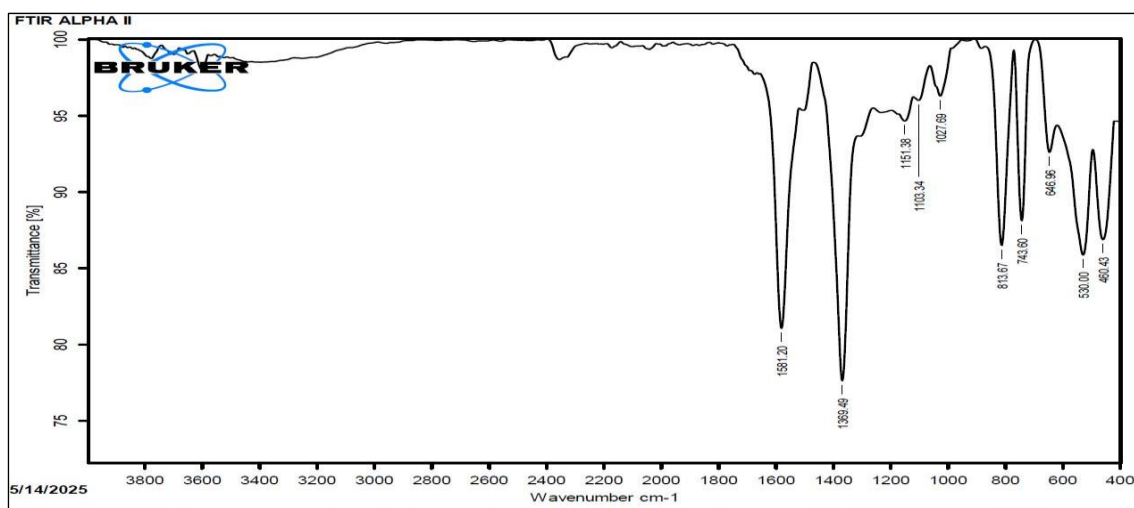
Where the  $R_e$  is represented recovery ratio and  $C_L$  is liberated concentration in solution by unit mg/L

### 3. Results and discussion

#### 3.1 FTIR characterization

The FTIR spectra of activated carbon (AC) and oxidized activated carbon (OAC) were examined to identify surface functional groups and verify the effectiveness of the oxidation treatment. The spectra displayed characteristic absorption bands at  $3063.7 \text{ cm}^{-1}$ , corresponding to aromatic C-H stretching vibrations, and at  $2970.6\text{-}2814.6 \text{ cm}^{-1}$ , attributed to aliphatic C-H stretching. A strong band observed at  $1677 \text{ cm}^{-1}$  is associated with C=C and/or C=O stretching vibrations, while the bands in the range of  $1572.4\text{-}1418.7 \text{ cm}^{-1}$  indicate the presence of benzene ring structures. Additional bands at  $1282.7$  and  $1110.2\text{-}1014.7 \text{ cm}^{-1}$  are assigned to C-O and C-O-C functional groups, respectively. Following nitric acid oxidation, the OAC spectrum exhibited a broad band centered around  $3400 \text{ cm}^{-1}$ , corresponding to -OH groups. The appearance and increased intensity of this band confirm the introduction of oxygen-containing functional groups, indicating successful surface oxidation and enhanced hydrophilicity of the modified carbon [36].

The FTIR spectrum of the oxide synthesized via the co-precipitation method exhibited strong absorption bands at  $498.14$  and  $624.7 \text{ cm}^{-1}$ , which are attributed to the Zn-O stretching vibrations. Additionally, prominent bands were observed at  $532.1$  and  $436.9 \text{ cm}^{-1}$ , along with a broad band centered at  $1050.2 \text{ cm}^{-1}$ , corresponding to the Fe-O and Fe-OOH vibrations. These results confirm the successful formation of zinc oxide and iron oxide phases [37]. Furthermore, the FTIR spectrum of the synthesized AC-Fe<sub>2</sub>O<sub>3</sub> and ZnO showed vibrations at  $460.4\text{-}646.9 \text{ cm}^{-1}$ , confirming the successful formation and loading of metal oxides onto the activated carbon surface. The decrease in the intensity of the carbon-related bands (-OH, C=O, C-O, and C-O-C) suggests a strong interaction between the oxide and the functional groups of the carbon matrix [21] as shown in Figure 1.



**Figure 1.** The FTIR result of prepared surface AC-ZnO and Fe<sub>2</sub>O<sub>3</sub> composite

### 3.2 BET analysis

The porous structure of the prepared composite plays a crucial role in the adsorption process, as it facilitates the diffusion and interaction of pollutant molecules through the pores, in addition to the contribution of the surface functional groups.

The Brunauer-Emmett-Teller (BET) method was employed to determine the specific surface area and pore characteristics of the synthesized carbon-based composite loaded with zinc and ferric oxide. The obtained results revealed that the prepared surface exhibited a specific surface area of 4.6932 m<sup>2</sup>/g, indicating the presence of a porous texture that enhances the adsorption capability of the composite materials (Table 1).

The low BET surface area of the composite is attributed to the following factors. The activated carbon was prepared from plastic waste and carbonized at a very low temperature (300 °C), which could not contribute to the formation of a well-defined microporous structure. Second, the loading of Fe<sub>2</sub>O<sub>3</sub> and ZnO nanoparticles into the carbon matrix partially blocked the pore entrance, thereby decreasing the surface area. Similar results of BET surface area reduction after incorporating metal oxides into the carbon matrix have been reported in previous studies. The low surface area of the composite did not affect the high adsorption performance, which shows that adsorption is not only dependent on the surface area but also on the surface functional groups, electrostatic attractions, and chemical affinity of interaction between the dye and the active sites [22, 38].

The BET surface area of the composite was relatively low (4.69 m<sup>2</sup>/g), but the adsorption capacity was very high. This implies that the physical surface area alone was not responsible for adsorption performance. In addition to the introduction of oxygen-containing functional groups (–OH and –COOH) by nitric acid oxidation, Fe<sub>2</sub>O<sub>3</sub> and ZnO nanoparticles added additional active sites. These surface functionalities can facilitate electrostatic attraction,

surface complexation, and possibly chemisorption of the dye molecules. The BET surface area is a nitrogen-accessible surface area of the dry material and may not accurately represent the available adsorption sites for large organic molecules in an aqueous environment. Similar reports of high adsorption capacity with a low BET surface area also exist for functionalized/metal-loaded carbon-based composites.

**Table 1.** The BET result of AC-ZnO-Fe<sub>2</sub>O<sub>3</sub>

Adsorption surface	Surface Area m <sup>2</sup> /g	Pore volume cm <sup>3</sup> /g	Pore diameter nm	Pore type
AC-ZnO- Fe <sub>2</sub> O <sub>3</sub>	4.6932	0.0308	2.4	Mesopores

### 3.3 X-ray powder diffraction

X-ray diffraction (XRD) was employed to determine the crystalline phase, crystal size, and degree of crystallinity of the prepared materials [39]. Distinct diffraction peaks indicate the material type and its crystallinity. For the activated carbon synthesized from plastic waste, characteristic peaks appeared at  $2\theta = 21.3^\circ, 23.05^\circ, 24.86^\circ, 25.59^\circ, 26.8^\circ, 32.08^\circ, \text{ and } 38.19^\circ$ , corresponding to the standard JCPDS card no. 41-1487. The prominent peak at  $2\theta = 25.59^\circ$  confirms the successful synthesis of activated carbon (Figure 2) [19].

The crystallite size (D) was estimated using the Debye-Scherrer equation [40]:

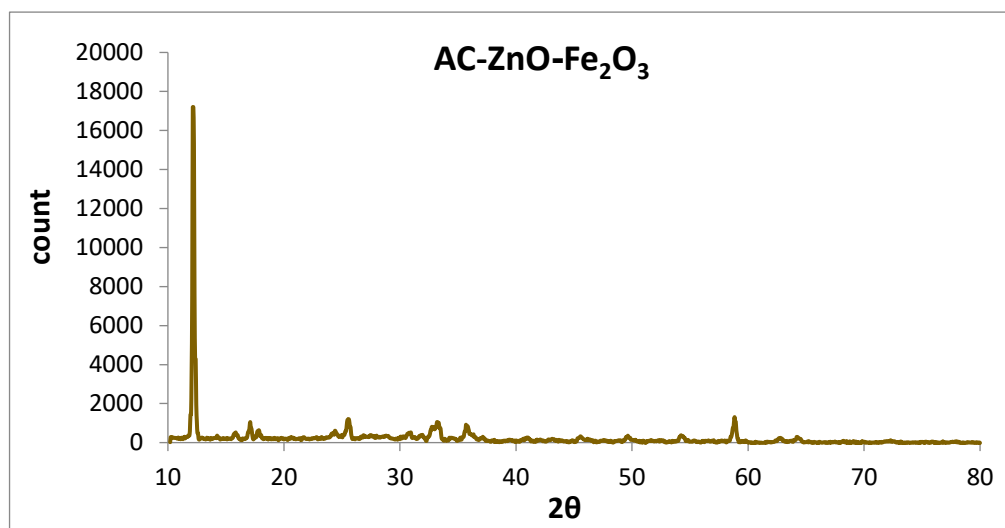
$$D = K\lambda / \beta \cos \theta$$

Where  $K=0.9$ ,  $\lambda = 1.54\text{A}$  (Cu  $\alpha$ ),  $\beta$  is the full width at half maximum of the peak, and  $\theta$  is the Bragg angle.

For zinc oxide (ZnO), the XRD pattern exhibited multiple peaks at  $2\theta = 23.03^\circ, 25.40^\circ, 34.77^\circ, 36.56^\circ, 38.35^\circ, 47.9^\circ, 56.86^\circ, 63.38^\circ, 66.9^\circ, 69.72^\circ, \text{ and } 77.8^\circ$ , which match the standard JCPDS card no. 36-1451 [41].

Similarly, the XRD pattern of ferric oxidize (Fe<sub>2</sub>O<sub>3</sub>) showed peaks at  $2\theta = 24.40^\circ, 25.42^\circ, 32.20^\circ, 33.54^\circ, 36.10^\circ, 38.27^\circ, 41.29^\circ, 49.98^\circ, 54.59^\circ, 58.09^\circ, 62.96^\circ, 64.45^\circ$ , consistent with JCPDS card no. 86-0550 confirming the successful co-precipitation synthesis of hematite

The XRD pattern of the AC-ZnO-Fe<sub>2</sub>O<sub>3</sub> composite displayed characteristic peaks of AC at  $2\theta = 12.18^\circ, 17.07^\circ, 25.59^\circ$  peaks corresponding to Fe<sub>2</sub>O<sub>3</sub> at  $2\theta = 24.37^\circ, 33.34^\circ, 54.28^\circ, 58.88^\circ$ . And peaks for ZnO at  $2\theta = 35.83^\circ, 45.6^\circ, 49.62^\circ, 62.78^\circ, 64.36^\circ, 72.18^\circ$ , confirming the successful hydrothermal loading of both metals oxidize onto the activated carbon surface [42].

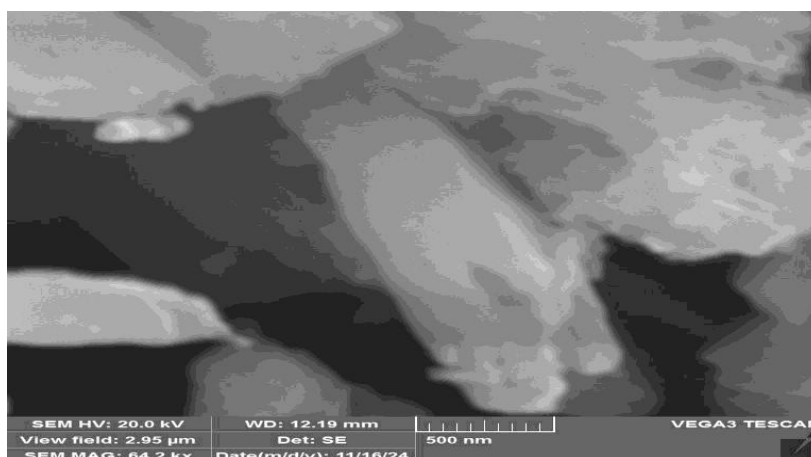


**Figure 2.** The XRD patterns of AC-ZnO -Fe<sub>2</sub>O<sub>3</sub> composite

### 3.4 Scanning Electron microscopy (SEM) Analysis

Scanning Electron microscopy (SEM) was employed to characterize the surface morphology, particle size, and porosity of the synthesized materials. The activated carbon (AC) derived from plastic waste exhibited a relatively homogeneous surface with well-developed and abundant porous channels, which are favorable for pollutant adsorption through interactions with surface functional groups. The average particle size, determined using Image-J software (version 1.46r), was 63.29nm, in good agreement with crystallite size estimated from X-ray diffraction (XRD) analysis.

The zinc oxide (ZnO) sample displayed a heterogeneous surface morphology characterized by particles with variable sizes and shapes and the presence of multiple surface pores, features that are expected to enhance adsorption efficiency. The average particle size was 68.79nm confirming that the material falls within the nanoscale range. In the case of iron oxide (Fe<sub>2</sub>O<sub>3</sub>), SEM micrographs revealed an irregular surface morphology with numerous pores, and an average particle size of 91.69nm, further supporting the successful synthesis of nanoscale particles. The AC-Fe<sub>2</sub>O<sub>3</sub>/ZnO composite exhibited diverse porous channels and cavities with a relatively uniform surface texture. The average particle size of this composite was 72.11nm, confirming the effective loading of metal oxides onto the activated carbon and the formation of an efficient nanoscale adsorbent (Figure 3).

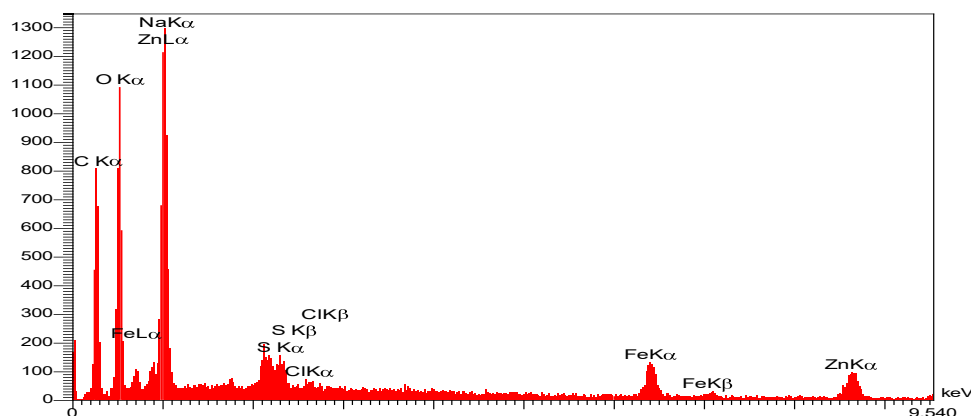


**Figure3.** The SEM morphology of AC-ZnO-Fe<sub>2</sub>O<sub>3</sub> composite

### 3.5 Energy Dispersive X-ray (EDX) Analysis

Energy-dispersive X-ray spectroscopy (EDX) was used to determine the elemental composition of the prepared materials at room temperature, using an accelerating voltage in the range of 0-80 keV, based on the emission of characteristic X-rays from each element. The EDX spectrum of activated carbon (AC) derived from plastic waste showed the presence of carbon and oxygen, indicating successful surface oxidation and an increased oxygen content on the carbon matrix. For zinc oxide (ZnO), the EDX analysis confirmed the presence of zinc and oxygen, verifying the successful formation of ZnO nanoparticles. The Fe<sub>2</sub>O<sub>3</sub> sample prepared via co-precipitation method exhibited iron and oxygen as the major elements, along with trace amounts of carbon, sodium, chlorine and silicon. These minor elements are attributed to residual precursors and impurities originating from the synthesis process.

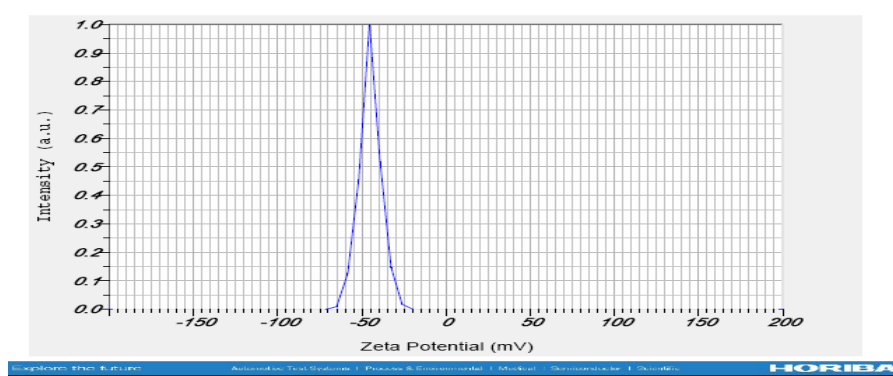
In the AC-Fe<sub>2</sub>O<sub>3</sub>/ZnO composite, the carbon-to-oxygen ratio was approximately 1.08, indicating the successful loading and distribution of ZnO-Fe<sub>2</sub>O<sub>3</sub> nanoparticles onto the activated carbon surface (Figure 4).



**Figure4.** The EDX result of AC-ZnO-Fe<sub>2</sub>O<sub>3</sub> composite

### 3.6 Zeta potential measurement

Zeta potential measurement is an essential technique to determine the surface charge of material and evaluate their stability in aqueous solution. The zeta potential of the high stability with a value of  $-42.2\text{mV}$ , indicating a negative charged surface. The value of  $-42.2\text{ mV}$  signifies that the surface is strongly negative charged and stable. As Eosin Blue and Fuchsin Acid are anionic dyes, electrostatic repulsion may be generated between the dye molecules and the adsorbent surface. Therefore, the adsorption process is not mainly governed by electrostatic attraction, but mainly by the non-electrostatic interactions, including  $\pi$ - $\pi$  stacking between the aromatic rings of the dyes and the graphitic structure of the activated carbon, hydrogen bonding between the functional groups of the dyes and the oxygen-containing groups ( $-\text{OH}$  and  $-\text{COOH}$ ) on the surface, as well as the dispersion forces and surface interactions of the  $\text{Fe}_2\text{O}_3$  and  $\text{ZnO}$  nanoparticles. Similar behavior has been reported for anionic dye adsorption onto negatively charged carbon-based composites [43] as shown in Figure 5.



**Figure5.** The zeta potential result of AC-ZnO-Fe<sub>2</sub>O<sub>3</sub> composite

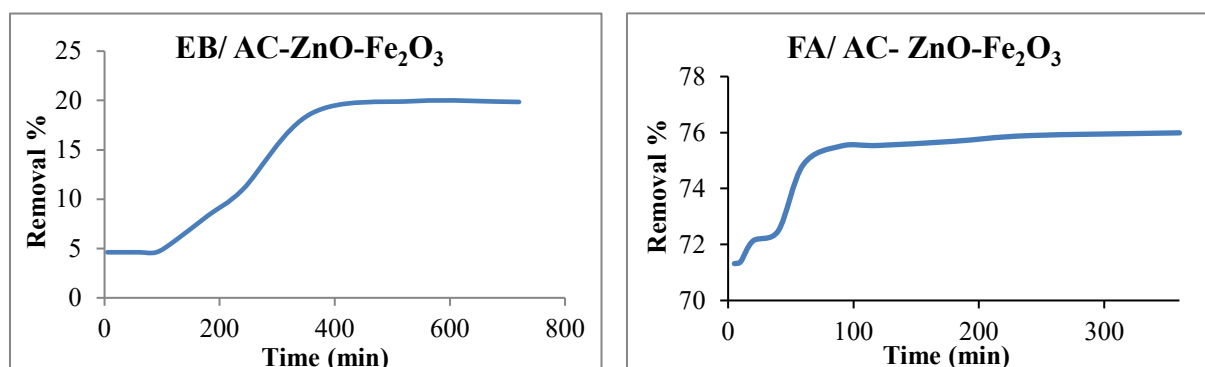
### 3.7 Adsorption

Four main parameters affecting the adsorption process were investigated namely: equilibrium contact time, solution pH, temperature, and agitation speed. Throughout these experiments, the adsorbent dosage (0.01g) and dye solution volume (10mL) were maintained constant. Eosin Blue and Fuchsin Acid were used as model pollutants at initial concentrations of 100mg/L and 200 mg/L, respectively.

#### 3.7.1 Effect of Equilibrium time

The effect of contact time on the adsorption of Eosin Blue and Fuchsin Acid onto AC-ZnO and Fe<sub>2</sub>O<sub>3</sub> was investigated at intervals ranging 5, 10, 20 ,40 ,60 ,90 ,120 ,180 ,240 ,360 ,540 ,720 min. Experiment was performed at room temperature, natural pH, and agitation speed of 150 rpm, with initial concentrations of 100 mg/L for EB and 200 mg/L for (FA). Results showed

that removal efficiency increased with contact time, reaching equilibrium at 360 min (18.71%) for EB and 90 min (75.52%) for FA likely due to differences in dye structure and adsorbent surface charge (Figures 6-7).

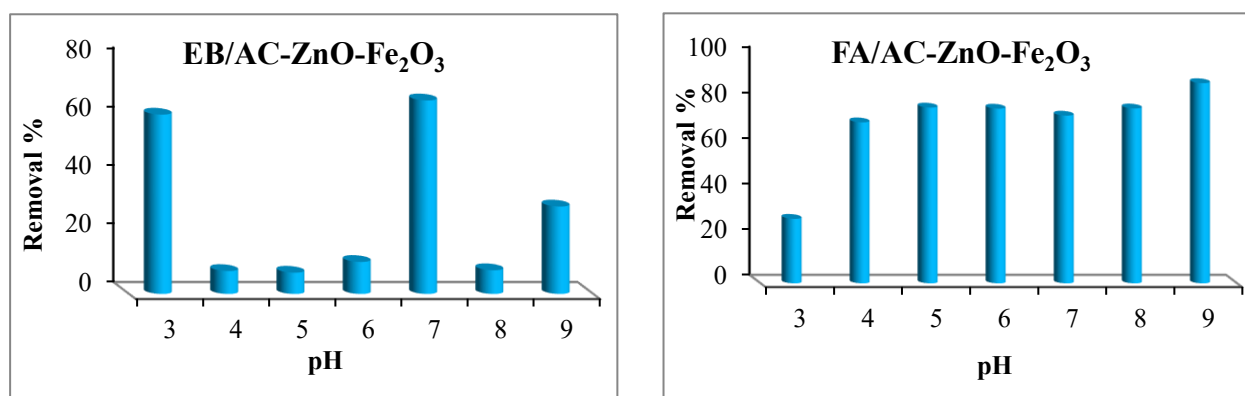


**Figure 7.** The effect of equilibrium time for FA **Figure 6.** The effect of equilibrium time for EB

### 3.7.2 Effect of pH

The effect of pH on the adsorption of Eosin blue and Fuchsin Acid onto AC-ZnO-Fe<sub>2</sub>O<sub>3</sub> was investigated by preparing dye solutions with fixed concentrations of 100 mg/L and 200 mg/L, respectively. Over a pH range of 3,4,5,6,7,8,9. experiment were conducted at room temperature, with agitation speed of 150rpm, using the previously determined equilibrium time for each dye.

The results revealed that the highest removal efficiency for Eosin blue was achieved at pH7 with 61.81%, followed by a gradual decrease at higher pH values, indicating electrostatic attraction between the different charged dye molecules and the adsorbent surface. In contrast, Fuchsin Acid exhibited a progressive increase in removal efficiency towards the alkaline range, reaching 87.73%, which can be attributed to the enhanced interaction between the different charge dye molecules and the adsorbent surface under basic condition (Figures 8-9).



**Figure 8.** The effect of pH for EB

**Figure 9.** The effect of pH for FA

### 3.7.3 Effect of Temperature

The effect of temperature variation on the adsorption efficiency of Eosin blue and Fuchsin Acid onto from their aqueous solution onto AC-ZnO-Fe<sub>2</sub>O<sub>3</sub> was investigated while keeping other parameters constant. Experiment was performed at different temperatures 25,35,45,55 and 65 °C, to determine the optimal adsorption temperature for both dyes, using initial concentrations of 100 mg/L for EB and 200 mg/L for FA, an agitation speed of 150 rpm, and the predetermined equilibrium time and pH. A fixed solution volume (10 mL) and adsorbent dose (0.01g) were used in all experiments. The results showed that the adsorption efficiency of both dyes was highest at 25°C, with removal percentage of 75.96% for EB and 88.10% for FA, as illustrated in Figures 10-11. The decrease in adsorption efficiency with increasing temperature indicates that the adsorption is overall exothermic, consistent with many dye–adsorbent systems where elevated temperature reduces interaction strength between dye molecules and surface-active sites.

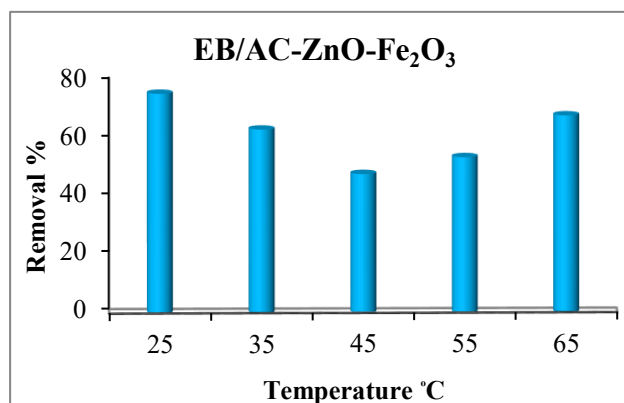


Figure 10. The effect of temperature for EB

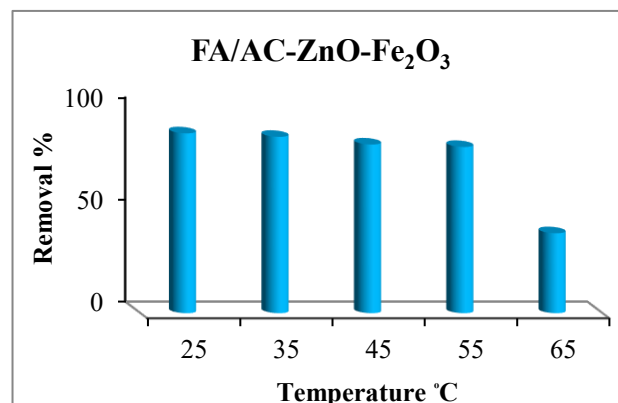
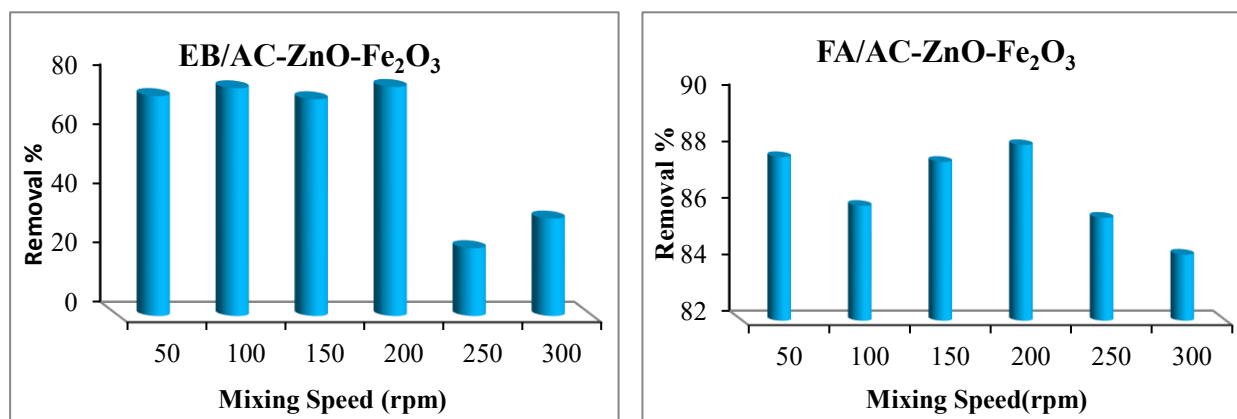


Figure 11. The effect of temperature for FA

### 3.7.4 Effect of agitation speed

The effect of agitation speed on the adsorption of Eosin blue and Fuchsin Acid from aqueous solution onto AC-ZnO-FeO<sub>3</sub> was studied while keeping the initial dye concentration constant at 100mg/L and 200 mg/L, respectively. A fixed dye solution volume (10mL) and adsorbent dose (0.01g) were used under the previously optimized conditions of pH, temperature, and contact time. The results showed that the highest adsorption efficiency was achieved at an agitation speed of 200 rpm, with removal percentage of 77.17% for Eosin blue and 88.21% for Fuchsin Acid, as illustrated in Figures12-13. The enhancement in adsorption with increasing stirring speed can be attributed to the improved diffusion rate of dye molecules toward the external surface of the adsorbent, reducing the boundary layer thickness and facilitating stronger interactions between the dye molecules and the active adsorption sites.



**Figure 12.** The effect of agitation speed for EB **Figure 13.** The effect of agitation speed for FA

After studying the four factors, the optimal conditions for the removal of Eosin blue and Fuchsin Acid were determined, as summarized in Table 2.

**Table 2.** The optimal condition for the adsorption process of Eosin Blue and Fuchsin Acid

Dye name	C <sub>0</sub> mg/L	Temperature °C	pH	Equilibrium time (min)	Agitation Speed rpm	Removal%
Eosin Blue (EB)	100	25	7	360	200	77.17
Fuchsin Acid (FA)	200	25	9	90	200	88.21

### 3.7.5 Adsorption Isotherms

The adsorption isotherms were evaluated to understand the adsorption behavior, surface characteristics, and interaction mechanisms between Eosin Blue and Fuchsin Acid molecules and the AC-Fe<sub>2</sub>O<sub>3</sub>/ZnO composite. In addition, the isotherm analysis was used to assess the nature of the adsorption process and the type of binding interactions involved. The equilibrium data were analyzed using the Langmuir and Freundlich isotherm models. The calculated parameters, summarized in Table 3 and presented in Figures 14-15, indicate that the adsorption of both dyes onto the modified activated carbon is favorable. The Freundlich heterogeneity factor ( $n > 1$ ) suggests favorable adsorption and reflects the heterogeneous surface interactions of the composite, with adsorption occurring on sites of different energies and affinities, consistent with a multilayer adsorption mechanism.

**Table3.** Freundlich constants for the adsorption of EB and FA

Dyes	Temperature °C	N	K <sub>f</sub>	R <sup>2</sup>
Eosin Blue (EB)	25	3.871828	31.34382	0.854
	35	3.189798	31.82782	0.890
	45	2.854542	31.19404	0.922
	55	2.85588	34.15859	0.900
Fuchsin Acid (FA)	25	2.107772	36.77251	0.954
	35	1.952858	32.19289	0.957
	45	1.82737	28.22733	0.958
	55	1.778619	27.04368	0.961

Furthermore, the correlation coefficients ( $R^2$ ) show that the adsorption of Fuchsin Acid is better described by the Freundlich model, indicating adsorption on a heterogeneous surface. In contrast, Eosin Blue exhibits a relatively lower fit to the Freundlich model, suggesting slight differences in adsorption behavior and interaction mechanisms between the two dyes. The favorability of adsorption was further evaluated using the Langmuir separation factor ( $R_L$ ), calculated as:

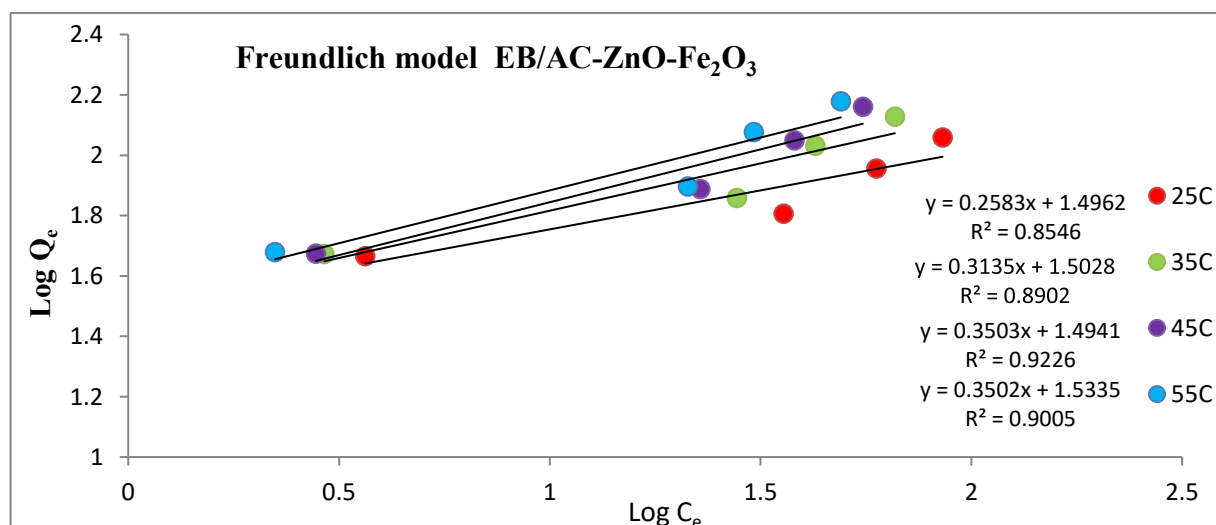
$$R_L = \frac{1}{1 + K_L C_0}$$

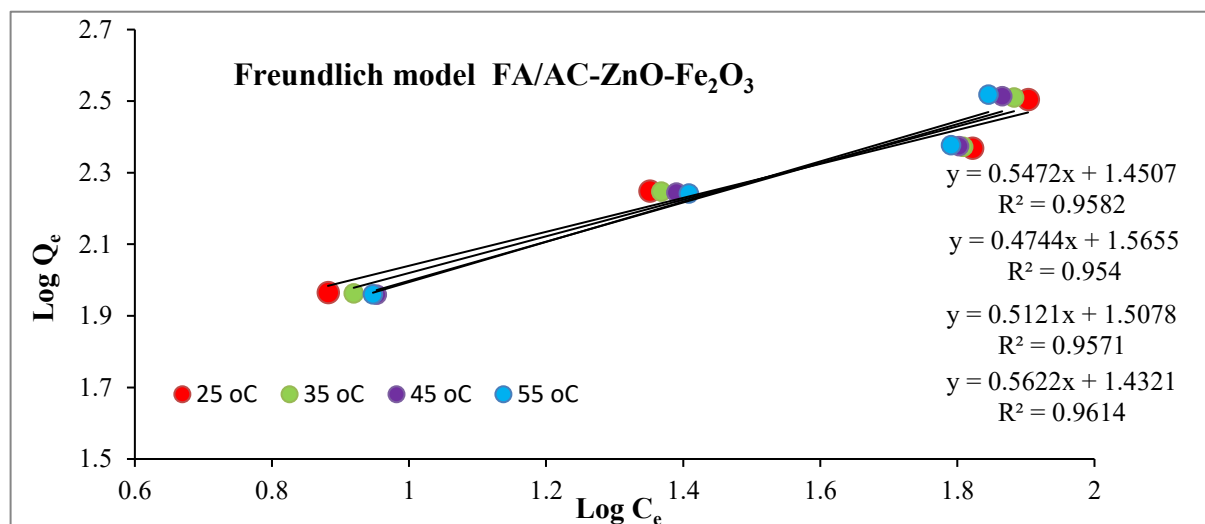
where:

$C_0$ : is the initial dye concentration (mg/L)

$K_L$ : is the Langmuir constant (L/mg)

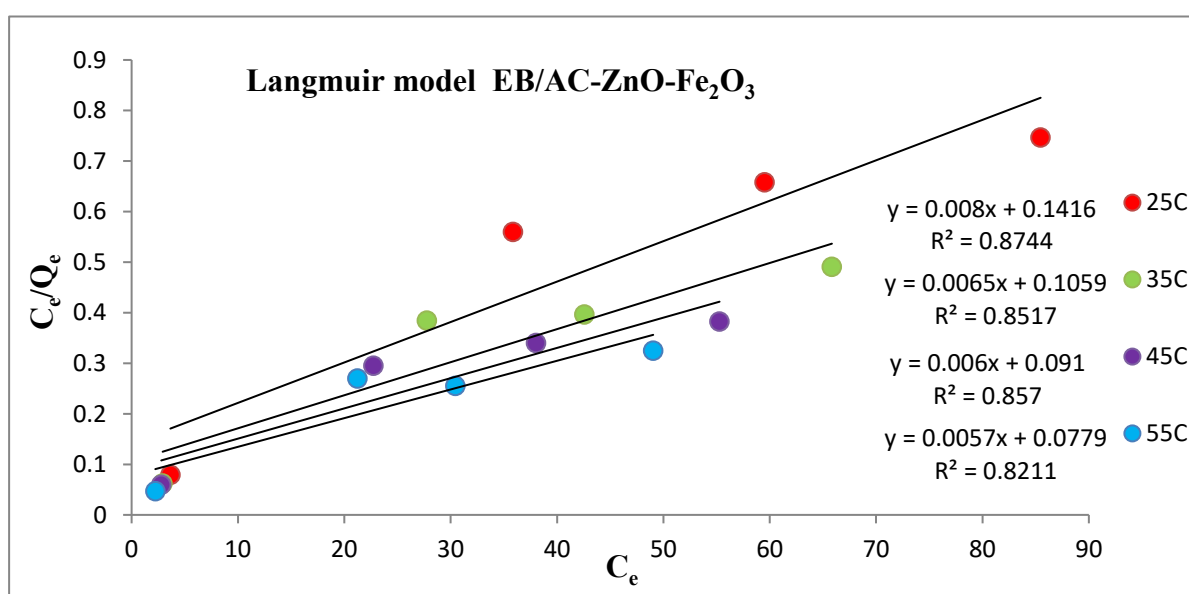
$R_L$ : is the dimensionless separation factor.

**Figure14.** Freundlich isotherm of Eosin Blue onto AC-ZnO-Fe<sub>2</sub>O<sub>3</sub> composite

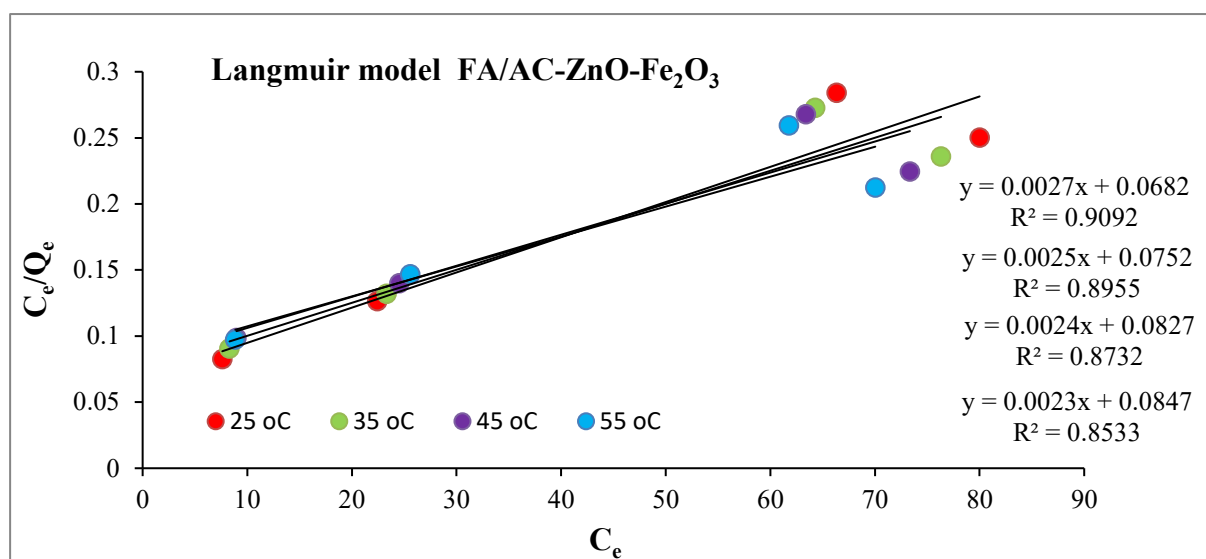


**Figure15.** Freundlich isotherm of Fuchsin Acid onto AC-ZnO-Fe<sub>2</sub>O<sub>3</sub> composite

The adsorption results fitted to the Langmuir model indicate the formation of a monolayer adsorption system, which can be attributed to the homogeneity of the adsorbent surface and the chemical nature of the dyes. Figures 16-17 present the adsorption behavior of Eosin blue and Fuchsin Acid on AC-Fe<sub>2</sub>O<sub>3</sub>&ZnO, showing a closer fit of Fuchsin Acid to the Langmuir model compared to Eosin blue (Table 4). The maximum adsorption capacity ( $Q_{max}$ ) was found to be 441.60 mg/g for Fuchsin Acid and 176.22 mg/g for Eosin Blue. Additionally, the separation factor ( $R_L$ ) values lie within the range  $0 < R_L < 1$ , confirming that the adsorption system is favorable and that interactions between the dye molecules and the adsorbent occur efficiently on a homogeneous monolayer surface



**Figure16.** Langmuir isotherm of Eosin Blue onto AC-ZnO-Fe<sub>2</sub>O<sub>3</sub> composite



**Figure17.** Langmuir isotherm of Fuchsin Acid onto AC-ZnO-Fe<sub>2</sub>O<sub>3</sub> composite

**Table 4.** Langmuir constants for the adsorption of EB and FA

Dyes	Temperature °C	$Q_{max}$ (mg/g)	$K_L$ (L/mg)	$R_L$	$R^2$
Eosin Blue (EB)	25	125.0258	0.056496	0.261452	0.982
	35	152.8847	0.061771	0.244585	0.988
	45	167.1524	0.065718	0.233323	0.968
	55	176.2279	0.072857	0.215386	0.613
Fuchsin Acid (FA)	25	375.4562	0.039032	0.203947	0.844
	35	400.3209	0.033216	0.231395	0.854
	45	425.4367	0.028436	0.260176	0.621
	55	441.6065	0.026725	0.272293	0.715

### 3.7.6 Thermodynamic study of the Adsorption process

Temperature plays a significant role in the adsorption process in solid-liquid systems and provides important information about the thermodynamic behavior of the system. Thermodynamic parameters such as Gibbs free energy ( $\Delta G^\circ$ ), enthalpy ( $\Delta H^\circ$ ), and entropy ( $\Delta S^\circ$ ) were evaluated to determine the spontaneity and nature of the adsorption [44]. The distribution coefficient (K), which represents the equilibrium between the amount of dye adsorbed on the solid and liquid phase and its concentration in the liquid phase, was calculated using:

$$K = \frac{Q_e}{C_e}$$

The Gibbs free energy change was obtained from the relation:

$$\Delta G^\circ = -RT \ln K$$

Negative  $\Delta G^\circ$  values indicate that adsorption is thermodynamically spontaneous under the studied conditions; however, spontaneity alone does not determine whether adsorption is purely physical or chemical. Mechanistic inference should be supported by the magnitude of thermodynamic parameters and complementary characterization. The relationship among thermodynamics functions follows the Gibbs equation [36]:

$$\Delta G^\circ = \Delta H^\circ - T\Delta S^\circ$$

The thermodynamic parameters ( $\Delta H^\circ$ ,  $\Delta S^\circ$ , and  $\Delta G^\circ$ ) were determined from the linear plots of  $\ln K$  versus  $1/T$ . The enthalpy change ( $\Delta H^\circ$ ) provides information about heat effect of the adsorption process, where a positive value indicates an endothermic process, while a negative value reflects exothermic behavior. The entropy change ( $\Delta S^\circ$ ) represents the degree of randomness at the solid-solution interface; positive  $\Delta S^\circ$  values suggest increased disorder and enhance affinity between the dye molecules and the adsorbent surface.

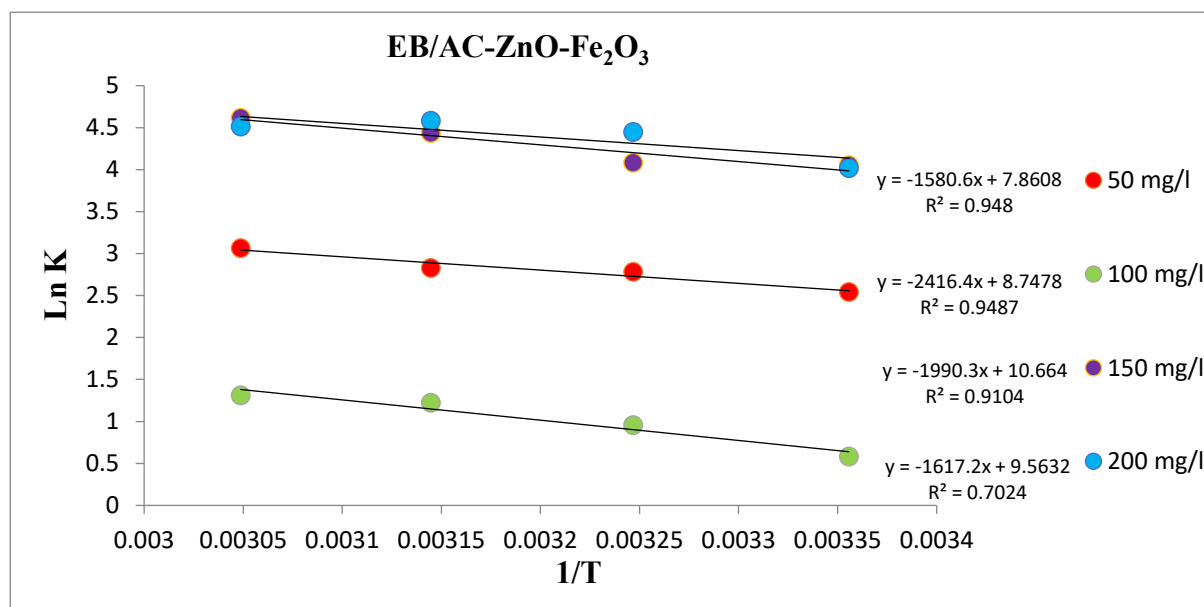
The calculated thermodynamic parameters for the adsorption of Eosin Blue (EB) and Fuchsin Acid (FA) onto the AC-Fe<sub>2</sub>O<sub>3</sub>/ZnO composite are presented in Tables 5-6 and illustrated in Figures 18-19. The negative  $\Delta G^\circ$  values obtained for both dyes at all studied temperatures confirm the spontaneous adsorption process. The relatively low magnitude of  $\Delta G^\circ$  further suggests that the adsorption is predominantly governed by physical interactions. For EB, positive values of  $\Delta H^\circ$  and  $\Delta S^\circ$  indicate that the adsorption process is endothermic and accompanied by an increase in randomness at the adsorbent-solution interface. In the case of FA, the negative  $\Delta G^\circ$  values also confirm spontaneous adsorption, while the positive  $\Delta S^\circ$  values reflect a strong affinity between the dye molecules and the composite surface.  $\Delta H^\circ$  for FA exhibited concentration-dependent behavior, showing exothermic characteristics at low concentrations (100-200 mg/L) and endothermic behavior at higher concentrations. These results indicate that the adsorption of both dyes is spontaneous and primary physical in nature, with temperature playing a significant role in the influencing the adsorption mechanism and the strength of interaction.

**Table 5.** Thermodynamic faction result of EB

C <sub>0</sub> mg/L	$\Delta H^\circ$ kJ/mol	$\Delta S^\circ$ J/mol/K	$\Delta G^\circ$ kJ/mol			
			298 °K	308 °K	318 °K	328 °K
50	13.1412	0.0653	-6.2974	-7.1222	-7.4813	-8.3578
100	20.0900	0.0727	-1.4400	-2.4493	-3.2319	-3.5752
150	16.5469	0.0886	-10.039	-10.451	-11.730	-12.588
200	13.4451	0.0795	-9.9507	-11.391	-12.113	-12.307

**Table 6.** Thermodynamic faction result of FA

C <sub>0</sub> mg/L	$\Delta H^\circ$ kJ/mol	$\Delta S^\circ$ J/mol/K	$\Delta G^\circ$ kJ/mol			
			298 °K	308 °K	318 °K	328 °K
100	-4.74964	0.004648	-6.18328	-6.15017	-6.12725	-6.35688
200	-4.08991	0.003503	-5.12415	-5.18576	-5.20123	-5.23452
300	2.361965	0.018417	-3.11935	-3.32738	-3.48213	-3.68122
400	4.390698	0.026256	-3.43449	-3.6996	-3.94925	-4.227

**Figure18.** Thermodynamic faction result of EB

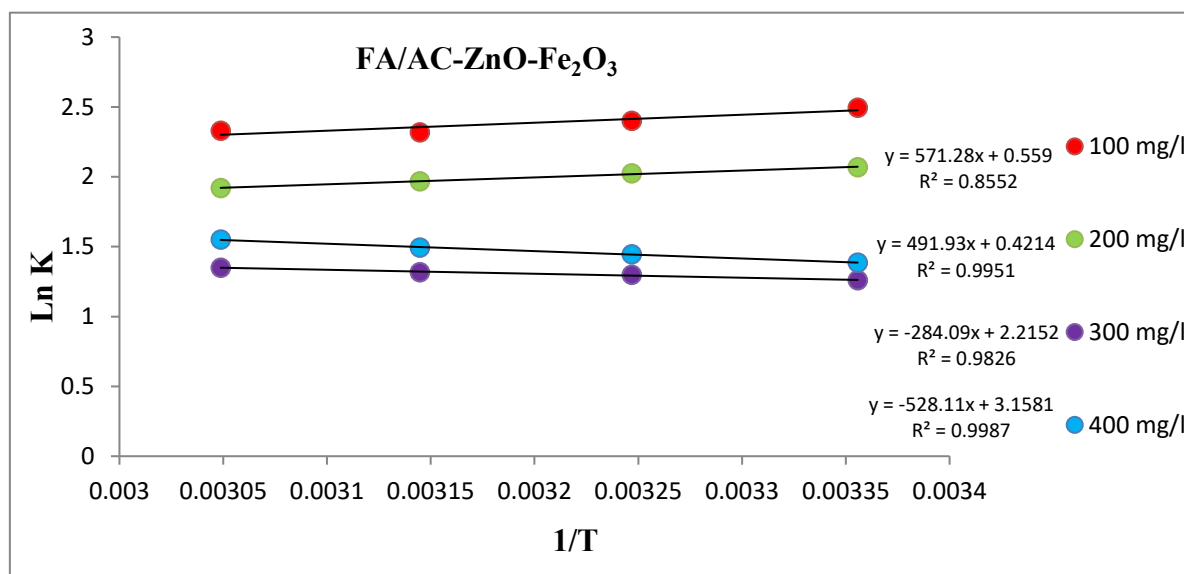
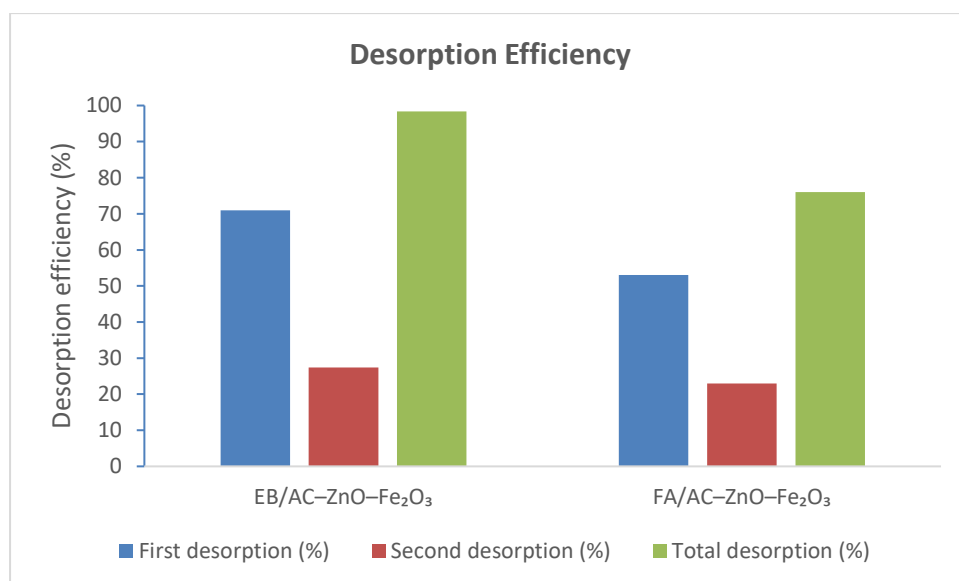


Figure19. Thermodynamic function result of FA

### 3.7.7 Recovery study

The desorption behavior of Eosin Blue (100mg/L) and Fuchsin Acid (200mg/L) from AC-Fe<sub>2</sub>O<sub>3</sub>/ZnO composite was evaluated using distilled water as a polar desorbing agent under continuous agitation. The desorption process was carried out in two successive stages to assess the regeneration efficiency of the adsorbent. For Eosin Blue, desorption efficiencies of 70.99% and 27.4% were achieved in the first and second stages, respectively, resulting in a total recovery of 98.46%. In contrast, Fuchsin Acid exhibited a lower cumulative recovery of 75.81% after two desorption cycles, indicating stronger interactions between the dye molecules and the adsorbent surface. The reduced recovery efficiency observed for FA may be attributed to stronger surface interactions, possible diffusion limitations within the pore structure, and the influence of molecular size and polarity on the desorption process (Figure20).



**Figure 20.** Desorption efficiency of Eosin Blue and Fuchsin Acid from AC-ZnO-Fe<sub>2</sub>O<sub>3</sub> composite

#### 4. Conclusions

The AC-ZnO-Fe<sub>2</sub>O<sub>3</sub> composite was successfully synthesized via a simple and environmentally sustainable approach using plastic waste derived activated carbon as a support for metal oxide nanoparticles. Comprehensive characterization and surface analyses (FTIR, EDX, XRD, BET, SEM and zeta potential) confirmed the successful formation of a stable composite with heterogeneous surface properties and enhanced surface functionality. The material exhibited high adsorption performance toward the studied anionic dyes, with a maximum adsorption capacity of 441.60 mg/g for Fuchsin Acid and 176.22 mg/g for Eosin Blue, indicating strong affinity and effective interaction between the adsorbent and dye molecules. Isotherm analysis revealed heterogeneous multilayer adsorption behavior, while thermodynamic results indicated that the adsorption process for both dyes was spontaneous and predominantly endothermic, with the exception of Fuchsin Acid at concentrations of 100 and 200 mg/L, where the process became exothermic. The integration of waste valorization with enhanced adsorption performance highlights the material as a low-cost and sustainable candidate for efficient treatment of dye-contaminated wastewater.

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## الامتزاز عالي الكفاءة والامتزاز المستدام للأصبغ الايونية باستخدام متراكب جديد من AC-ZnO-Fe<sub>2</sub>O<sub>3</sub>

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### المستخلص

هدفت هذه الدراسة إلى تحضير وتوصيف مادة نانوية مركبة جديدة ماصة عالية الكفاءة مكون من الكربون المنشط المحمل بأكاسيد الزنك والحديد (AC-ZnO-Fe<sub>2</sub>O<sub>3</sub>) مطورة من الكربون المنشط المستخلص من النفايات البلاستيكية وجسيمات أكاسيد المعادن النانوية، وذلك لإزالة صبغة الايوسين الأزرق وحامض الفوشين من المحاليل المائية بشكل مستدام. تم توصيف الخصائص الفيزيائية والكيميائية والبنية والمورفولوجية للمادة المحضرة باستخدام تقنيات FTIR و XRD و FE-SEM و EDX و BET وتحليل جهد زيتا. وأكدت نتائج تجارب الامتزاز تحقق اعلى كفاءة ازالة بلغت 77.17% و 88.21% لكل من صبغتي الايوسين الازرق وحامض الفوشين على التوالي. وكشفت دراسات الاتزان ان بيانات الامتزاز تتوافق بشكل أفضل مع نموذج فريندلش مما يشير إلى حدوث امتزاز متعدد الطبقات على سطح غير متجانس. وأشار التحليل الديناميكي الحراري إلى الطبيعة التلقائية لعملية الامتزاز (ΔG سالبة)، التي كانت ماصة للحرارة في الغالب. كما اظهرت تجارب الاسترجاع امكانية اعادة استخدام المادة بكفاءة عالية، إذ بلغت نسب الاسترجاع 98.46% لصبغة الايوسين الأزرق و 75.81% لصبغة حامض الفوشين باستخدام الماء كذيب. ويستنتج أن المركب النانوي المحضر يعد مادة ماصة فعالة ومنخفضة التكلفة ومستدامة بيئيًا لمعالجة المياه العادمة الصناعية الملوثة بالأصبغ وإعادة استخدامها.

**الكلمات المفتاحية:** الكاربون المنشط، الايوسين الازرق، الفوشين، اوكسيد الزنك، الملوثات العضوية، الاسترجاع.